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North American influence on tropospheric ozone and the effects of recent emission reductions: Constraints from ICARTT observations

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[1] We use observations from the International Consortium for Atmospheric Research on Transport and Transformation (ICARTT) campaign over eastern North America in summer 2004, interpreted with a global 3-D model of tropospheric chemistry (GEOS-Chem), to improve and update estimates of North American influence on global tropospheric ozone and the effect of recent U.S. anthropogenic reductions on surface ozone pollution. We find that the 50% decrease in U.S. stationary NO_x sources since 1999 has decreased mean U.S. boundary layer ozone concentrations by 6–8 ppbv in the southeast and 4–6 ppbv in the Midwest. The observed $d\text{O}_3/d\text{CO}$ molar enhancement ratio in the U.S. boundary layer during ICARTT was $0.46 \text{ mol mol}^{-1}$, larger than the range of 0.3–0.4 from studies in the early 1990s, possibly reflecting the decrease in the NO_x/CO emission ratio as well as an increase in the ozone production efficiency per unit NO_x . North American NO_x emissions during summer 2004 as constrained by the ICARTT observations (0.72 Tg N fossil fuel, 0.11 Tg N biomass burning, 0.28 Tg N lightning for 1 July to 15 August) enhanced the hemispheric tropospheric ozone burden by 12.4%, with comparable contributions from fossil fuel and lightning (5–6%), but only 1% from biomass burning emissions despite 2004 being a record fire year over Alaska and western Canada.

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1. Introduction

[2] Ozone (O_3) is produced in the troposphere by photochemical oxidation of methane (CH_4), carbon monoxide (CO) and nonmethane volatile organic compounds (NMVOCs) in the presence of nitrogen oxides ($\text{NO}_x = \text{NO} + \text{NO}_2$). Anthropogenic emissions of NO_x and methane have caused a large global increase of tropospheric ozone over the past century [Marenco *et al.*, 1994; Wang and Jacob, 1998] with consequences for the greenhouse effect

[Intergovernmental Panel on Climate Change, 2007] and surface air pollution [Economic Commission for Europe, 2007]. A number of studies have used $d\text{O}_3/d\text{CO}$ concentration enhancement ratios in continental outflow, where CO is a long-lived tracer of combustion, to estimate continental anthropogenic influences to ozone on the global scale [Mauzerall *et al.*, 1998; Pochanart *et al.*, 1999; Zhang *et al.*, 2006] including from North America [Chin *et al.*, 1994; Parrish *et al.*, 1993, 1998]. We use here observations from the International Consortium for Atmospheric Research on Transport and Transformation (ICARTT) aircraft campaign over eastern North America in summer 2004 [Singh *et al.*, 2006; Fehsenfeld *et al.*, 2006], interpreted with a global 3-D model of tropospheric chemistry (GEOS-Chem) to improve and update estimates of North American influence on global tropospheric ozone including both anthropogenic and natural sources. Stationary NO_x and mobile CO sources in the United States have been decreasing rapidly over the past decade [Frost *et al.*, 2006; Kim *et al.*, 2006; Parrish, 2006; Hudman *et al.*, 2007, 2008]. The ICARTT campaign provides constraints on these decreases as well as on natural sources of ozone precursors from lightning and fires.

[3] CO, a product of incomplete combustion, has a mean atmospheric lifetime of 2 months against oxidation by the hydroxyl radical (OH), its main atmospheric sink. It is a tracer for continental outflow, and the ozone-CO relationship is a marker for ozone production in a polluted air mass.

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Previous observations in U.S. boundary layer air and North American outflow in summer have shown consistently strong correlations with enhancement ratios $dO_3/dCO = 0.3\text{--}0.4 \text{ mol mol}^{-1}$ [Chin *et al.*, 1994; Mao and Talbot, 2004; Parrish *et al.*, 1993, 1998]. Although CO per se is not a limiting precursor for ozone production, combined information on dO_3/dCO and CO emissions enables estimate of anthropogenic ozone export [Parrish *et al.*, 1993; Chin *et al.*, 1994]. Here, we examine recent changes in this relationship due to recent emission reductions.

[4] Hudman *et al.* [2007, 2008] previously applied GEOS-Chem to the ICARTT data to derive constraints on U.S. emissions of CO and NO_x from combustion and lightning. They found that the ICARTT data implied a 50% reduction in anthropogenic NO_x emissions from stationary sources (power plant and industry) relative to the National Emission Inventory of the U.S. Environmental Protection Agency (EPA), constructed for 1999 (NEI 99), consistent with 1999–2004 regulations of stationary NO_x sources. Additionally, the ICARTT data implied a summertime U.S. anthropogenic CO source 60% lower than specified in the NEI 99 inventory and weaker than the biogenic CO source from the oxidation of VOCs. A larger lightning source than expected from models was found in the upper troposphere over the eastern United States [Martin *et al.*, 2006; Hudman *et al.*, 2007]. This is consistent with recent studies suggesting higher NO_x yields per flash from lightning at northern midlatitudes than in the tropics [Huntrieser *et al.*, 2008; Ott *et al.*, 2007].

[5] The summer of 2004 was one of the strongest fire seasons on record for Alaska and western Canada [Turquety *et al.*, 2007]. Elevated levels of CO and ozone associated with these fires were observed as far as the central Atlantic [Cook *et al.*, 2007; Val Martín *et al.*, 2006] and western Europe [Real *et al.*, 2007]. Here, we examine the quantitative impacts of the fires on the hemispheric-scale ozone budget.

2. ICARTT

[6] ICARTT was a coordinated multi-aircraft and ground-based atmospheric chemistry field program over eastern North America and the North Atlantic in July–August 2004. Two important components were the NOAA Intercontinental Transport and Chemical Transformation (ITCT 2k4) and the NASA Intercontinental Transport Experiment–North America, Phase A (INTEX-A) campaigns. The ITCT2k4 campaign [Fehsenfeld *et al.*, 2006] took place 3 July to 15 August over the NW Atlantic and the NE United States out of Portsmouth, New Hampshire. It used a WP-3D aircraft (ceiling 6 km). The INTEX-A campaign [Singh *et al.*, 2006] took place 29 June to 14 August over the central and eastern United States, and the North Atlantic, from bases at Edwards Air Force Base (California), St. Louis (Missouri), and Portsmouth (New Hampshire). It used a DC-8 aircraft (ceiling 12 km).

[7] We use here the measurements of CO, ozone, NO₂, and NO taken aboard both aircraft (the DC-8 NO data are those from Pennsylvania State University [Ren *et al.*, 2008]). Intercomparisons between the two aircraft showed agreement to within the combined accuracy of the measurements for CO, ozone, and NO₂ (NO data were not com-

pared) (G. Chen, NASA Langley, personal communication, 2006, <http://www-air.larc.nasa.gov/missions/intexna/meas-comparison.htm>). We use the ITCT2k4 and INTEX-A measurements as a single data set in our analysis. We also make use of other data sets collected on the aircraft including peroxyacetylnitrate (PAN), nitric acid (HNO₃), hydrogen cyanide (HCN), and acetonitrile (CH₃CN). Details on the aircraft data are given by Fehsenfeld *et al.* [2006] and Singh *et al.* [2006].

[8] Surface measurements were taken during ITCT2k4 from Chebogue Point on the southern tip of Nova Scotia (44°N, 66°W) to observe outflow from the northeastern United States [Fehsenfeld *et al.*, 2006]. Millet *et al.* [2006b] found that CO, ozone, and aerosol mass concentrations were elevated respectively by 30%, 56%, and more than 300% at Chebogue Point during U.S. outflow periods. Here we use the CO and ozone data averaged over 30-min intervals.

3. Model Description

[9] We simulate the ICARTT observations with the GEOS-Chem global 3-D model of tropospheric chemistry (version 7.02; <http://www.as.harvard.edu/chemistry/trop/geos/>) driven by assimilated meteorological observations from the Goddard Earth Observing System (GEOS-4) of the NASA Global Modeling and Assimilation Office (GMAO). The model is applied to a global simulation of ozone–NO_x–VOC–aerosol chemistry. A general description of GEOS-Chem is given by Bey *et al.* [2001] and a specific description of the coupled oxidant–aerosol simulation as used here is given by Hudman *et al.* [2007]. The horizontal resolution of the model is 2° × 2.5°. There are 30 vertical layers including 12 below 10 km and 5 below 2 km. The simulation was conducted for 15 months starting from chemical climatology; the 13-month spin-up time effectively removes the influence of initial conditions.

[10] For comparison with observations, the model is sampled along the aircraft flight tracks for the flight times. Fine-scale fresh pollution and biomass burning plumes must be excluded from this comparison because of intrinsically large model errors associated with plume displacement and dilution over a 2° × 2.5° grid box. We diagnose fresh pollution plumes in the observations by NO_x/NO_y > 0.4 mol mol^{−1} (here, NO_y = NO_x + PAN + HNO₃). When NO_y is not available we diagnose fresh pollution plumes by NO₂ > 4 ppbv. We also remove fresh biomass burning plumes as diagnosed by HCN > 500 pptv or CH₃CN > 225 pptv [de Gouw *et al.*, 2006]. When filtered, all species over that time period are removed. Several of the WP-3D flights were targeted at urban and power plant plume characterization [Fehsenfeld *et al.*, 2006] and we further exclude those data. Overall these filters exclude 7% and 21% of the DC-8 and WP-3D data, respectively.

[11] Global anthropogenic emissions in the model are as described by Park *et al.* [2004]. Emissions for the contiguous United States are summarized in Table 1. Anthropogenic emissions are from NEI 99 with modifications described by Hudman *et al.* [2007, 2008], including a generalized 50% decrease in NO_x emissions from power plants and industry reflecting 1999–2004 reductions [Frost *et al.*, 2006], and a 60% decrease in CO emissions reflecting both a decreasing trend and a likely overestimate in the

Table 1. CO and NO_x Sources in the Contiguous United States for 1 July to 15 August 2004

Source Type	CO (Tg)	NO _x ^a (g N)
Fuel and industry	6.4 ^b	0.62 ^c
Vegetation	9.1 ^d	-
Soil	-	0.12 ^e
Biomass Burning ^f	0.16	0.003
Aircraft	-	0.021
Lightning ^g	-	0.27
Total	15.7	1.03

^aNO_x emissions are as in the work by *Hudman et al.* [2007].

^bIncluding a secondary source of 1.8 Tg from oxidation of anthropogenic VOCs, calculated for the purposes of this table by assuming VOC-specific CO yields from *Duncan et al.* [2007]. The anthropogenic emission of CO (4.6 Tg CO) represents a 60% reduction from the NEI 99 inventory to match the ICARTT observations [*Hudman et al.*, 2008].

^cIncluding 0.28 Tg N from transport, 0.17 Tg N from power generation and industry reflecting a 50% reduction from the NEI 99 inventory to match the ICARTT observations [*Hudman et al.*, 2007], and 0.17 Tg N from other sources.

^dIncluding 6.7 Tg from isoprene oxidation and 2.4 Tg from oxidation of other biogenic VOCs (acetone, monoterpenes, = C₃ alkenes). For the purposes of this table, we assume a 2.3 molar yield of HCHO from isoprene oxidation as constrained by the ICARTT observations [*Millet et al.*, 2006a] and further assume 100% conversion of HCHO to CO. For the other biogenic VOCs we use the CO yields from *Duncan et al.* [2007].

^eIncluding 0.03 Tg from fertilizer application.

^fNorth American fires during ICARTT were mainly outside the contiguous United States; Alaskan and Canadian fires produced 19 Tg CO and 0.10 Tg N during this period.

^gContiguous United States and coastal waters (130°–70°W, 25°–50°N).

original NEI 99 inventory [*Hudman et al.*, 2008]. Biogenic emissions of isoprene and monoterpenes are from the GEIA inventory [*Guenther et al.*, 1995]. Monoterpenes are accounted as sources of CO with an instantaneous yield of 0.2 CO per atom C [*Duncan et al.*, 2007]. Biogenic alkene (C ≥ 3) emissions are assumed to be 10% of isoprene on a per molecule basis. Biogenic acetone emissions are as in the work by *Jacob et al.* [2002].

[12] We increase the U.S. lightning NO_x source by a factor of four relative to the standard GEOS-Chem scheme (i.e., to 500 mol/flash) to match constraints from the upper tropospheric NO_x aircraft data and lightning observations from the National Lightning Detection Network (NLDN) [*Hudman et al.*, 2007].

[13] Boreal fire emissions for CO are from the ICARTT daily inventory of *Turquety et al.* [2007], resulting in 19 Tg CO for the 1 July to 15 August period. There have been very few measurements of emission factors for NO_x from boreal fires. On the basis of a summary of available observations, *Lapina et al.* [2008] estimate molar NO_x/CO emission ratios of 3×10^{-3} mol/mol for smoldering and 28×10^{-3} mol/mol for flaming combustion. Following *Kasischke and Bruhwiler* [2003], *Turquety et al.* [2007] assume 50% flaming and 50% smoldering combustion, except for peat (assumed to be 100% smoldering), which results in an overall mean NO_x/CO emission ratio of 0.011, which we use here. Fire emissions during ICARTT were frequently lofted to the free troposphere [*de Gouw et al.*, 2006]. While injection heights are uncertain, we assume following *Turquety et al.* [2007] that 40% of fire emissions were released at the surface and 60% in the free troposphere with uniform vertical distribution by mass.

[14] The boundary ventilation rate in the model has been validated through the successful simulation of the observed shapes of the mean vertical profiles for propane [*Hudman et al.*, 2008], acetylene [*Xiao et al.*, 2007], and formaldehyde [*Millet et al.*, 2006a]. Additionally, NO_x emissions vented to the free troposphere as NO_y (observed = $16 \pm 10\%$, modeled = $14 \pm 9\%$) and its partitioning is well captured [*Hudman et al.*, 2007]. Because of its short lifetime in the boundary layer (<1 day), surface NO_x concentrations are primarily governed by local emissions and chemical loss [*Li et al.*, 2004, Table 3]. After applying the 1.7 correction factor to DC-8 OH measurements (as discussed by *Ren et al.* [2008]), the OH comparison is unbiased [*Hudman et al.*, 2007], suggesting the NO_x lifetime in the model is well represented.

4. Effect of U.S. NO_x Emission Reductions on Surface Ozone

[15] Figure 1 shows observed and simulated afternoon ozone concentrations at 0–1.5 km altitude averaged over the GEOS-Chem grid boxes. For comparison with observations, the model was sampled along the aircraft flight tracks for the flight times and then averaged. Cool temperatures, cloudiness, and frequent frontal passages made summer 2004 a mild ozone season over the eastern United States [*Gilliland et al.*, 2008]. Mean observed concentrations were typically 45–60 ppbv. High values observed offshore reflect aged urban pollution plumes deliberately sampled by the WP-3D aircraft in stratified air [*Neuman et al.*, 2006] and should not be viewed as representative mean of the grid box. The simulation using NEI 99 is too high over the United States by 5–15 ppbv ($R^2 = 0.38$).

[16] *Frost et al.* [2006] determined from stack sampling that power plant emissions of NO_x decreased by 50% between 1999 and 2003, resulting in an 11% reduction in total anthropogenic NO_x emissions during the summer ozone season. Previously, *Hudman et al.* [2007] showed that ICARTT NO_x observations imply a 22% reduction in anthropogenic NO_x emissions relative to NEI 99. This NO_x reduction improves the ozone simulation in the south and Midwest, as shown in Figure 1, though it only slightly improves the overall model-to-observed ozone correlation ($R^2 = 0.41$).

[17] The NO_x emission reductions (Figure 2) are largest in the Midwest, reflecting the distribution of power plants. The largest resulting ozone decreases in the model (up to 8 ppbv) are in the southeast (Figure 2); the Midwest shows weaker decreases (4–6 ppbv). *Kim et al.* [2006] examined the effect of power plant NO_x emission reductions, following *Frost et al.* [2006], on simulated ozone levels by using the WRF-Chem regional model at 27×27 km² resolution applied for summer 2004. They found a similar spatial distribution of ozone changes, with local concentrations decreasing by as much as 7 ppbv and greater ozone reductions per unit NO_x in the south.

[18] The spatial variability in ozone response to NO_x emission reductions can be explained by variability in the ozone production efficiency (OPE), defined as the number of ozone molecules produced (*P*) per molecule of NO_x oxidized (*L*), PO_3/LNO_x [*Liu et al.*, 1987]. OPE decreases with increasing NO_x and increases with increasing isoprene and UV flux [*Liu et al.*, 1987; *Hirsch et al.*, 1996], and is

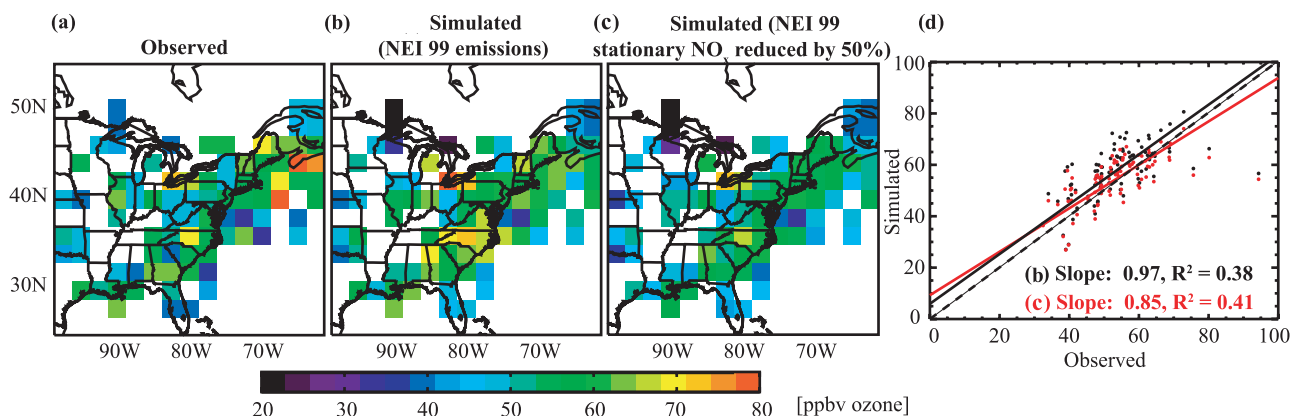


Figure 1. Mean afternoon (1100–1700 LT) ozone concentrations at 0–1.5 km altitude during ICARTT. (a) Observations are compared to (b) the model simulation with NEI 99 emissions and (c) the simulation with NO_x stationary sources decreased by 50%. (d) A scatterplot of this gridded comparison and $y = x$ line (dashed).

thus higher in the southeast than in the northeast. We find in the GEOS-Chem simulation that the Midwest boundary layer has OPE values in the range 2.5–3.5, whereas the southeast has values in the range 4–5.5, consistent with the simulated pattern of ozone decreases.

5. Ozone-CO Relationship in the Boundary Layer

[19] Figure 3 shows simulated versus observed ozone-CO correlations for the boundary layer aircraft flight tracks (0–1.5 km altitude) in the afternoon (1100–1700 LT), for photochemically aged air as diagnosed by $\text{NO}_x/\text{NO}_y < 0.3 \text{ mol mol}^{-1}$ following *Chin et al.* [1994]. Also shown are correlations at Chebogue Point for the W-SE wind sector which sampled North American outflow [*Millet et al.*, 2006b]. Model results are from the standard simulation

including reduced NO_x and CO emissions relative to NEI 99 (Table 1). The $d\text{O}_3/d\text{CO}$ enhancement ratio (as defined by the slope of the reduced-major-axis regression line, allowing for errors on both variables and 95% confidence intervals, determined by nonparametric bootstrap resampling) is $0.46 \pm 0.02 \text{ mol mol}^{-1}$ ($R^2 = 0.56$) in the aircraft observations and $0.36 \pm 0.02 \text{ mol mol}^{-1}$ ($R^2 = 0.23$) in the model sampled along the aircraft flight tracks. The discrepancy is caused principally by a population of model points in the southeast with relatively high ozone (40–60 ppb) and low CO, reflecting an ozone overestimate in inflow from the Gulf of Mexico as previously discussed by *Fiore et al.* [2002]. The observed $d\text{O}_3/d\text{CO}$ enhancement ratio at Chebogue Point is 0.41 ± 0.03 ($R^2 = 0.47$), consistent though slightly lower than the aircraft data, likely due to surface deposition. The full ensemble of Chebogue Point data (all

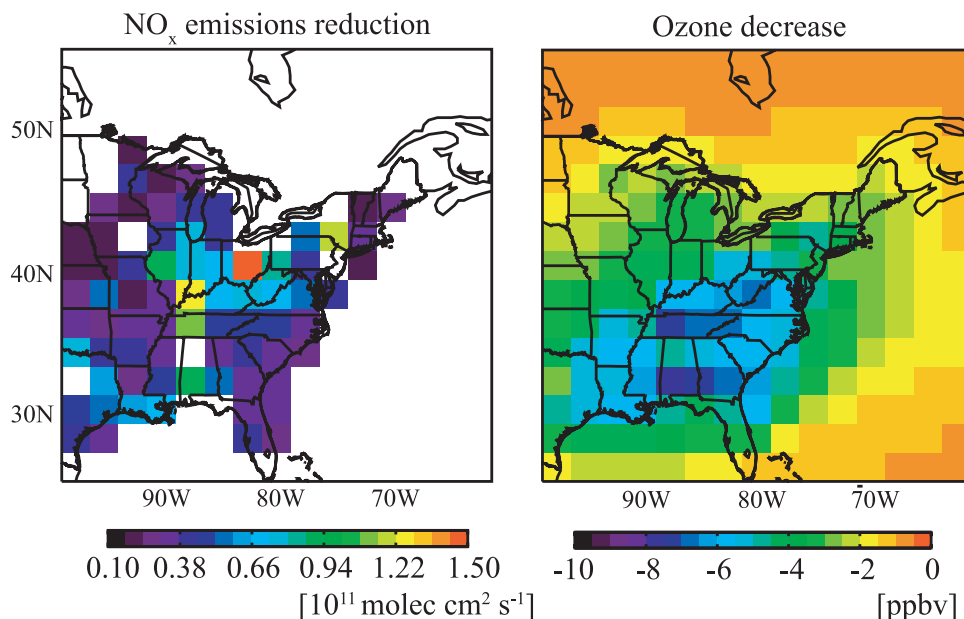


Figure 2. (left) U.S. NO_x emission reductions from 1999 to 2004 due to 50% reduction in stationary NO_x sources and (right) resulting decrease of mean afternoon ozone concentrations at 0–1.5 km altitude (1100–1700 LT) simulated for the ICARTT period.

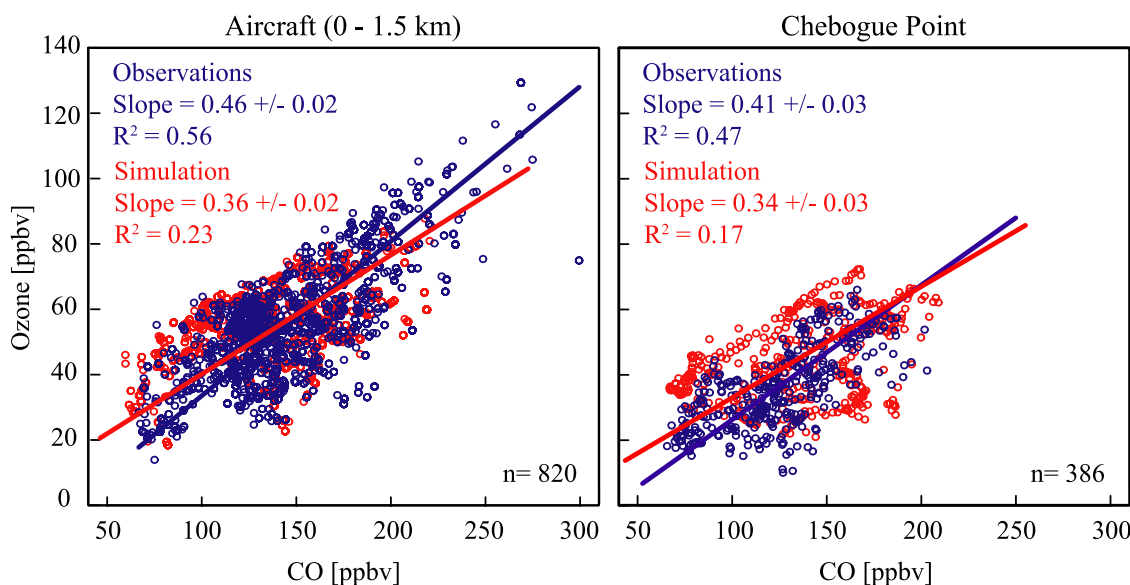


Figure 3. Ozone-CO correlations in the boundary layer during ICARTT: (left) aircraft flight tracks and (right) Chebogue Point. Observations (blue) are compared to model results (red). Flight tracks are for the afternoon (1100–1700 LT) in the boundary layer and east of 100°W. Chebogue Point data are for the subset with observed winds from SE-W (120°–270°W). Slopes are from a reduced-major-axis regression.

wind sectors) has $dO_3/dCO = 0.47 \pm 0.03$, but the R^2 degrades to 0.26. The model is somewhat too low with dO_3/dCO of 0.34 ± 0.03 ($R^2 = 0.17$), and this again appears to be due to an excessive ozone background.

[20] Previous observations of the summertime dO_3/dCO enhancement ratios in the United States had consistently found values of 0.3–0.4. *Chin et al.* [1994] report enhancement ratios at three eastern U.S. sites of 0.28–0.32 ($R^2 = 0.43$ –0.78) for 1988–1992. *Parrish et al.* [1998] report ratios at four sites in Atlantic Canada of 0.28–0.40 ($R^2 = 0.48$ –0.68) for 1991–1994. The abnormally cool summer in 2004 suppressed ozone production, suggesting observed changes in dO_3/dCO would be a lower bound. The higher values seen during ICARTT could reflect decadal changes in U.S. CO and NO_x emissions. *Mao and Talbot* [2004], however, found a slightly lower dO_3/dCO of 0.37 ($R^2 = 0.77$) over New Hampshire during southwesterly flow for the summers 2001–2003.

[21] The dO_3/dCO enhancement ratio in North American outflow can be interpreted as the number of ozone molecules produced per molecule of CO emitted. If we assume that NO_x is in steady state between emission and loss ($E_{NO_x} = L_{NO_x}$), we can further interpret the dO_3/dCO enhancement ratio as the product of the OPE (P_{O_3}/L_{NO_x}) and the molar NO_x/CO source ratio ($E_{N/C} = E_{NO_x}/E_{CO}$), the former taken from the model and the latter constrained by ICARTT observations [*Hudman et al.*, 2007, 2008]:

$$\frac{dO_3}{dCO} \approx OPE \times E_{N/C} \quad (1)$$

Consider the evolution of $E_{N/C}$ over the northeast United States (EPA regions 1–3, <http://www.epa.gov/epahome/locate2.htm>) for the period 1994–2004. The 1 July to 15 August 2004 anthropogenic emissions for the northeast

were 1.2 Tg CO and 0.10 Tg N, to which we must add 0.87 Tg CO from biogenic emissions. *Parrish* [2006] estimates urban anthropogenic CO emissions decreased by 4.6% a^{−1} during the period 1987–2002. Assuming that this trend continued through 2004 and that the CO source from NMVOC oxidation (mainly biogenic) remained constant, we estimate a total CO source decrease of 23% over 1994–2004. NO_x emissions from powerplants remained roughly constant until 2000, at which time reductions in power sources began to be phased in, resulting in a 50% decrease by summer 2004 [*Frost et al.*, 2006]. *Hudman et al.* [2007] found that to match ICARTT NO_x observations, large industry must similarly be reduced by 50%, resulting in a 22% decrease in NO_x emissions nationwide and a 15% decrease in the northeast. The decadal change in the mobile source of NO_x is uncertain (estimates range from slight decreases to slight increases) and thus we assume constant mobile emissions over the period [*Parrish*, 2006; *Martin et al.*, 2006; *Hudman et al.*, 2007; *Gilliland et al.*, 2008].

[22] We thus estimate a 10% increase in the molar NO_x/CO source ratio $E_{N/C}$ over 1994–2004. We further find in the model that the NO_x emission reduction relative to the NEI 99 inventory caused a mean increase in the OPE over the northeast of 9%. The product of the trends in the NO_x/CO source ratio and in the OPE would thus be expected to yield a 19% increase in the dO_3/dCO enhancement ratio, consistent with the trend in the observed ratio as described above.

6. Hemispheric Ozone Enhancements From North American Anthropogenic, Biomass Burning, and Lightning Emissions

[23] The ICARTT observations provided strong constraints on North American sources of NO_x from anthropo-

Table 2. North American Enhancements to the Summertime Northern Hemisphere Burden of Tropospheric Ozone^a

North American Source	NO _x Emission ^b (Tg N)	Ozone Production Efficiency	Hemispheric Ozone Enhancement ^c (Tg, %)
Lightning	0.28	32	9.0 (5.1%)
Biomass burning	0.11	18	2.0 (1.1%)
Fossil fuel	0.72	15	10.9 (6.2%)
Total	1.11	20	21.9 (12.4%)

^aObtained by difference between the standard simulation and a sensitivity simulation with the corresponding source shut off, and with sufficient (11 months) initialization in each case. Here and elsewhere, North America is defined as Canada and the continental United States. Values are for the 1 July to 15 August 2004 ICARTT period.

^bFor the 1 July to 15 August 2004 period.

^cTotal Northern Hemisphere ozone burden is 175.4 Tg in the standard GEOS-Chem simulation.

genic emissions, lightning, and biomass burning, and implied large corrections to prior emission estimates (Table 1 and Hudman *et al.* [2007]). Table 2 gives the NO_x emission totals from each source for North America, defined here as the U.S. and Canada. We use here this information to quantify the North American contribution to tropospheric ozone on the hemispheric scale by conducting sensitivity simulations with anthropogenic, biomass burning, or lightning emissions over North America shut off. The associated ozone enhancement is diagnosed by difference with results from the standard simulation. We do not examine the sensitivity to the soil source as it is small compared to the anthropogenic source (Table 1). Each simulation is initialized with an 11-month spin-up.

[24] Figure 4 compares simulated and observed mean vertical distributions of ozone for the ensemble of DC-8 flights along with the simulated anthropogenic, biomass burning, and lightning enhancements. The total North American enhancement amounts to about 30 ppbv in the boundary layer and 20 ppbv in the free troposphere. The boundary layer enhancement is mainly anthropogenic, but the free troposphere has equal enhancements from anthropogenic and lightning NO_x. Biomass burning adds 1 ppbv from the surface to 4 km and less above.

[25] Figure 5 shows the ozone enhancements from the different North American NO_x sources on the scale of the Northern Hemisphere at different altitudes, and Table 2 summarizes the corresponding hemispheric budgets. Li *et al.* [2005] found that a semipermanent upper level anticyclone above the southern United States in summer allows ozone buildup in the upper troposphere by trapping convectively lifted precursors and lightning NO_x. Cooper *et al.* [2006] confirmed the resulting ozone maximum by analysis of ozonesonde data during ICARTT, and found an associated 11–14 ppbv contribution to ozone from lightning. We find a similar 10–17 ppbv enhancement in the upper troposphere over the Gulf Coast and see it stretching downwind across the subtropical Atlantic to Europe. North American lightning contributes 5.1% of the total Northern Hemisphere ozone burden for the ICARTT period, comparable to the North American anthropogenic ozone burden (Table 2). Even though the North American lightning NO_x source is less than half of the fossil fuel source, its OPE is much higher (Table 2) [also see Wu *et al.*, 2007].

[26] The mean simulated North American anthropogenic ozone enhancement is highest in the eastern United States in surface air but shifts northeast to the North Atlantic in the free troposphere, reflecting the ventilation of the U.S.

boundary layer by frontal passages [Fuelberg *et al.*, 2007]. In the upper troposphere, the anthropogenic enhancement extends to higher latitudes than that from lightning. We find that North American anthropogenic emissions enhance the tropospheric ozone burden in the Northern Hemisphere by 6.1% (Table 2). Enhancements in surface air over western Europe are 3–6 ppbv, consistent with previous estimates by Li *et al.* [2002].

[27] Biomass burning ozone enhancements are highest in boreal North America, reflecting the distribution of fires in the summer 2004, and extend to the Arctic, where they are comparable to fossil fuel influence in the lower and middle troposphere, though not in the upper troposphere. North American biomass burning emissions during the summer 2004 enhanced hemispheric ozone burden by 1% (2 Tg).

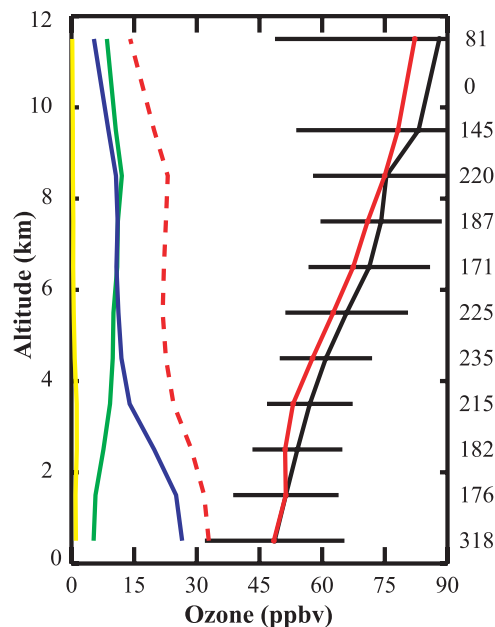


Figure 4. Mean vertical profiles of ozone over eastern North America during ICARTT. Observations from the DC-8 aircraft (black) are compared to model results sampled along the aircraft flight tracks (red). Model enhancements from North American biomass burning (yellow), lightning (green), anthropogenic emissions (blue) and all three sources combined (dashed red) are shown. Horizontal bars are standard deviations on the observations. Number of gridded observations per 1-km bin is shown.

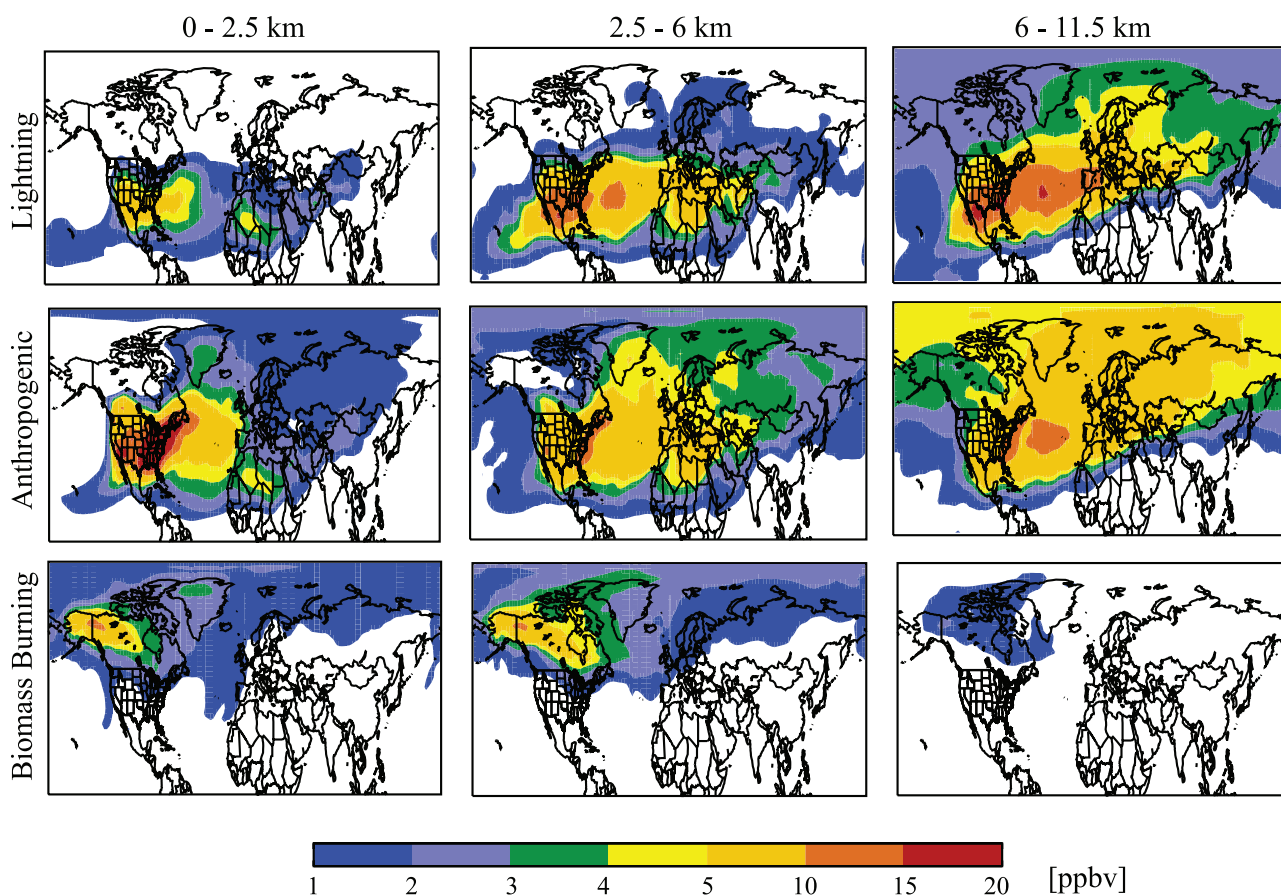


Figure 5. Simulated ozone enhancements from North American lightning, anthropogenic, and biomass burning emissions at 0–2.5, 2.5–6, and 6–11.5 km altitude as diagnosed by sensitivity simulation with individual sources turned off as described in section 6. Values are means for 1 July to 15 August 2004.

Observations of fire plumes at the PICO-NARE station, in the central Atlantic, showed mean dO_3/dCO in aged fire plumes of 0.2 mol mol^{-1} [Val Martín *et al.*, 2006], which if applied to the 19 Tg CO inventory used here would correspond to a 3% (6 Tg) ozone enhancement. Such an enhancement is consistent with a modeling study by Pfister *et al.* [2006] which used a molar NO_x/CO emission ratio of 0.27, 2.5x higher than used here. Using molar NO_x/CO emission ratios from 0.007 to 0.32, Cook *et al.* [2007] found corresponding increases in hemispheric ozone burdens during ICARTT of 1–3% (2–6 Tg). As discussed in Section 3, observational estimates of NO_x/CO emission ratios are closer to the lower end of this range. It thus appears that boreal fire emissions have a negligible influence on summertime ozone concentrations on the hemispheric scale.

7. Conclusions

[28] The ICARTT study in summer 2004 provided extensive observations of ozone, CO, and NO_y species over the eastern United States and western North Atlantic, from the surface to 12 km altitude. These data offer new information to better constrain and understand North American influence on tropospheric ozone over the scale of the Northern Hemisphere. Hudman *et al.* [2007] used the ICARTT observations of NO_y species from two aircraft (NASA DC-8 and NOAA WP-3D), together with simulations with

the GEOS-Chem chemical transport model (CTM), to improve estimates of U.S. fossil fuel and lightning NO_x sources. Here we used the ICARTT data together with GEOS-Chem to examine the effect of recent NO_x emission reductions [Frost *et al.*, 2006; Hudman *et al.*, 2007] on surface ozone in the United States, to quantify and interpret ozone–CO correlations seen in pollution outflow, and to assess the implications for the effect of natural and anthropogenic North American emissions on Northern Hemispheric ozone.

[29] Hudman *et al.* [2007] found that the ICARTT boundary layer NO_x data are consistent with a 50% summertime decrease in power plant and industry NO_x emissions in the eastern United States between 1999 and 2004 [Frost *et al.*, 2006] yielding a 22% decrease in the overall anthropogenic NO_x source in that season. Applying this decrease in the GEOS-Chem model reduces simulated ozone concentrations by 6–8 ppbv in the southeast and 4–6 ppbv in the Midwest, improving the comparison to observations. Even though the NO_x emission reduction is largest in the Midwest, the ozone decrease is greater in the southeast due to higher ozone production efficiency per unit NO_x (OPE).

[30] Aircraft data during ICARTT show a consistent dO_3/dCO enhancement ratio of $0.46 \text{ mol mol}^{-1}$ over the eastern United States and in U.S. outflow. Previous studies

using data from the early 1990s had found a consistent summertime enhancement ratio of 0.3–0.4, lower than in ICARTT. We show that this increase can be explained by a combination of OPE increase (9%) due to decreasing NO_x emissions and increase in the NO_x/CO source ratio (10%) due to even more rapid decrease of CO emissions.

[31] The ICARTT data offer a unique ensemble of constraints on NO_x emissions from fossil fuel combustion, lightning, and biomass burning in North America during summer 2004. We used the NO_x emissions derived from these constraints to better quantify North American influence from these different sources on tropospheric ozone over the scale of the Northern Hemisphere. We find that North American NO_x emissions in summer 2004 increased the hemispheric tropospheric ozone burden by 12%, with comparable contributions from fossil fuel and lightning (5–6%), but only 1% from biomass burning.

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